

Environmental Studies Faculty Publications

Environmental Studies

1-23-2020

Sex Ratio is Variable and Increasingly Male Biased at Two Colonies of Magellanic Penguins

Natasha J. Gownaris Gettysburg College

Pablo García Borboroglu University of Washington - Seattle Campus

P. Dee Boersma University of Washington - Seattle Campus

Follow this and additional works at: https://cupola.gettysburg.edu/esfac

Part of the Animal Sciences Commons, and the Environmental Sciences Commons

Share feedback about the accessibility of this item.

Recommended Citation

Gownaris, N.J., García Borboroglu, P., & Boersma, P.D. (2020). Sex Ratio is Variable and Increasingly Male Biased at Two Colonies of Magellanic Penguins. *Ecology*, *101*(3): e02939.

This is the author's version of the work. This publication appears in Gettysburg College's institutional repository by permission of the copyright owner for personal use, not for redistribution. Cupola permanent link: https://cupola.gettysburg.edu/esfac/109

This open access article is brought to you by The Cupola: Scholarship at Gettysburg College. It has been accepted for inclusion by an authorized administrator of The Cupola. For more information, please contact cupola@gettysburg.edu.

Sex Ratio is Variable and Increasingly Male Biased at Two Colonies of Magellanic Penguins

Abstract

Sex ratios are commonly skewed and variable in wild populations, but few studies track temporal trends in this demographic parameter. We examined variation in the operational sex ratio at two protected and declining breeding colonies of Magellanic Penguins (Spheniscus magellanicus) in Chubut, Argentina. Penguins from the two colonies, separated by 105 km, migrate north in the non-breeding season and have overlapping distributions at sea. Conditions during the non-breeding season can impact long-term trends in operational sex ratio (i.e., through sex-specific survival) and interannual variation in operational sex ratio (i.e., through sex-specific breeding decisions). We found an increasingly male-biased operational sex ratio at the two disparate colonies of Magellanic Penguins, which may contribute to continued population decline. We also found that the two colonies showed synchronous interannual variation in operational sex ratio, driven by variation in the number of females present each year. This pattern may be linked to sex-specific overwintering effects that cause females to skip breeding, i.e., to remain at sea rather than returning to the colony to breed, more often than males. Contrary to our predictions, colony-wide reproductive success was not lower in years with a more male-biased operational sex ratio. We did find that males showed more evidence of fighting and were less likely to pair when the operational sex ratio was more male biased. Our results highlight an indirect mechanism through which variation in the operational sex ratio can influence populations, through a higher incidence of fighting among the less abundant sex. Because biased sex ratios can reduce the size of the breeding population and influence rates of conflict, tracking operational sex ratio is critical for conservation.

Keywords

penguins, seabirds, sex ratio, conservation, breeding

Disciplines

Animal Sciences | Environmental Sciences

1	Running Head: Male-Biased Sex Ratios in Magellanic penguins
2	
3	Sex ratio is variable and increasingly male-biased at two colonies of Magellanic penguins
4	
5	Gownaris, N.J. ^{1*} , García Borboroglu, P. ^{2,3} , Boersma, P.D. ^{2,3,4}
6	
7	¹ Environmental Studies Department, Gettysburg College, Gettysburg, PA, USA, 17325
8	² Department of Biology and Center for Ecosystem Sentinels, University of Washington, Seattle,
9	WA, USA, 98103
10	³ Global Penguin Society, Puerto Madryn, Argentina and San Francisco, CA, USA, 94107
11	⁴ CESIMAR CCT Cenpat-CONICET, 9120 Puerto Madryn, Chubut, AR
12	
13	
14	*Natasha Gownaris, Corresponding Author
15	ngownaris@gmail.com
16	Environmental Studies Department, Gettysburg College, Gettysburg, PA 17325
17	
18	
19	
20	
21	
22	Keywords

23 mate competition, sex ratio, seabirds, skipped breeding, surveys, interannual variation

24 Abstract

Sex ratios are commonly skewed and variable in wild populations, but few studies track 25 temporal trends in this demographic parameter. We examined variation in the operational sex 26 ratio at two protected and declining breeding colonies of Magellanic penguins (Spheniscus 27 *magellanicus*) in Chubut, Argentina. Penguins from the two colonies, separated by 105 km, 28 29 migrate north in the non-breeding period and having overlapping distributions at sea. Conditions during the non-breeding season can impact long-term trends (i.e., through sex-specific survival) 30 31 and interannual variation in the operational sex ratio (i.e., through sex-specific breeding 32 decisions).

We found an increasingly male-biased operational sex ratio at the two disparate colonies of Magellanic penguins, which may contribute to continued population decline. We also found that the two colonies showed synchronous interannual variation in the operational sex ratio, driven by variation in the number of females present each year. This pattern may be linked to sex-specific overwintering effects that cause females to skip breeding, i.e., to remain at sea rather than returning to the colony to breed, more often than males.

Contrary to our predictions, colony-wide reproductive success was not lower in years with a more male-biased operational sex ratio. We did find that males showed more evidence of fighting and were less likely to pair when the operational sex ratio was more male-biased. Our results highlight an indirect mechanism through which variation in the operational sex ratio can influence populations, through a higher incidence of fighting among the less abundant sex. Because biased sex ratios can reduce the size of the breeding population and influence rates of conflict, tracking operational sex ratio is critical for conservation.

47 Introduction

Evolutionary theory suggests that population sex ratios should be balanced and stable 48 49 (Fisher 1930). However, sex ratios are often skewed and variable across space and time in wild populations (e.g., Le Gaillard et al. 2005, Ewen et al. 2011, Székely et al. 2014ab, Morrison et al. 50 2016, Pöysä et al. 2019). Skewed adult sex ratios (ASRs) result from biased sex ratios at birth or 51 52 from sex-specific mortality or permanent emigration (Székely et al. 2014b). While ASR is based solely on the number of breeding individuals of each sex, the operational sex ratio (OSR) 53 represents the number of individuals of each sex that are available to mate (Ancona et al. 2017). 54 55 Thus, OSR also depends on sex-specific breeding decisions (Ancona et al. 2017). Skewed ASRs can reduce population growth rate by reducing the size of the effective 56 breeding population (Jenouvrier et al. 2010, Eberhart-Phillips et al. 2017, Gownaris and Boersma 57 2019) and via subtler impacts, such as those on breeding behavior. Male-biased populations may 58 59 show an increase in aggression of males towards each other (Lee et al. 2017, Öst et al. 2018), 60 towards females (Ewen et al. 2011), or towards offspring (Stokes and Boersma 2000). Mating systems are also impacted by male-biased sex ratios, which have been linked to reduced parental 61 cooperation (Eberhart-Phillips et al. 2018), increased rates of extrapair mating and divorce 62 63 (Székely et al. 2014a, Heinsohn et al. 2019), and a higher incidence of brood desertion (Székely et al. 2006). Though often studied in the context of ASR, male-biased OSR may have similar 64 65 impacts on aggression, mating systems, and offspring survival (e.g., Rankin et al. 2011). 66 Adult sex ratios are commonly male-biased in bird populations (Donald 2007), most 67 often driven by female-biased mortality in juveniles and adults (Liker et al. 2005). While many 68 terrestrial bird species can be sexed visually, seabirds often have low or no sexual dimorphism 69 (Schreiber and Burger 2001). As a result, sex-specific survival is rarely considered in these

70	populations (Gownaris and Boersma 2019). Because seabirds are long-lived species, sex biases									
71	in survival can accumulate with age and time, resulting in increasingly skewed ASRs and OSRs									
72	over the age of a cohort or colony (Ainley and DeMaster 1980, Lambertucci et al. 2012,									
73	Gownaris and Boersma 2019).									
74	Additionally, individuals of many seabird species may remain at sea instead of returning									
75	to their colony to breed (i.e., skip breeding) if they are in poor body condition or environmental									
76	conditions are unfavorable (in red-footed booby Sula sula: Cubaynes et al. 2011; in common									
77	guillemots Uria aalge: Reed et al. 2015). These decisions are consistent with the "Prudent Parent									
78	Hypothesis", which suggests that long-lived species should prioritize survival over current									
79	reproduction (Drent and Daan 1980). Non-breeding conditions often have sex-specific carry-over									
80	effects on individual body condition and breeding performance (Salton et al. 2015, Saino et al.									
81	2017, Rebstock and Boersma 2018), so breeding decisions may also be sex-specific. Sex-specific									
82	skipped breeding rates could therefore drive high interannual variation in OSR.									
83	We examine temporal trends a variation in the OSR of Magellanic penguins (Spheniscus									
84	magellanicus) at two breeding colonies separated by 105 km in Chubut, Argentina (Appendix									
85	Text S1). Magellanic penguins, like many other seabirds, are long-lived, socially monogamous,									
86	and have obligate biparental care (Boersma et al. 2013). Punta Tombo (PT) covers an area									
87	approximately 20 times that of Cabo Dos Bahias (CDB) and has a much larger breeding									
88	population (Pozzi et al. 2015; Figure 1). The colonies' foraging areas overlap during egg									
89	incubation and to some extent during chick rearing (Boersma et al. 2009). After the breeding									
90	season, penguins from both colonies migrate north and overwinter at sea between Northern									
91	Argentina and Southern Brazil (Stokes et al 2014).									

Mortality of juvenile and adult Magellanic penguins is highest during the non-breeding 92 season and is female-biased (Vanstreels et al. 2011, Vanstreels et al. 2013). Female-biased 93 94 mortality has contributed to reduced population growth rate and a male-biased ASR at PT (Gownaris and Boersma 2019). Furthermore, when conditions in the species' non-breeding 95 habitat cause reduced food availability (i.e., a stronger Rio de la Plata plume), females but not 96 97 males return to the breeding colony in lower body condition (Rebstock and Boersma 2018). Oceanographic conditions may therefore affect both long-term trends in OSR (i.e., through 98 99 accumulative effects of sex-biased mortality) and interannual trends in OSR in Magellanic 100 penguins (i.e., through sex-specific effects of oceanographic conditions on breeding decisions). Because Magellanic penguins that breed at PT and CBD overlap in their nonbreeding 101 range, we expected the two colonies to show similar patterns of OSR (Hypothesis 1). At both 102 colonies we predicted: an increasingly male-biased OSR (Prediction H1.1) and synchronous 103 104 interannual variation in OSR (Prediction H1.2). We expected variation in OSR to show a 105 stronger correlation to variation in the number of females present than to variation in the number of males present (Prediction H1.3). We also tested the hypothesis that OSR is a driver of 106 breeding behavior and success at PT (Hypothesis 2). Specifically, we predicted that years with 107 108 more male-biased OSR would be marked by lower reproductive success (Prediction H2.1) due to 109 two non-mutually exclusive possibilities: 1) females returning in these years are in worse body 110 condition due to low food availability during the non-breeding season (e.g., Rebstock and 111 Boersma 2018), 2) rates of male aggression towards eggs or chicks are higher when OSR is more male-biased (e.g., Stokes and Boersma 2000). Lastly, we predicted that male Magellanic 112 penguins would fight more frequently (Prediction H2.2) and be less likely to secure a mate 113 114 (Prediction H2.3) in years with a more male-biased OSR.

115 Methods

116 <u>Colony Monitoring: Annual Surveys</u>

117 We conducted annual surveys at PT and CDB during settlement and incubation (October; started in 1987 and 1992 breeding seasons at PT and CDB, respectively) and during late-chick 118 rearing (January; started in 1986 and 1993 breeding seasons at PT and CDB, respectively) in 119 120 most years. Permanently staked survey plots are separated by 33.3 m at CDB and by 100 m at PT. We based our analyses on a subset of plots surveyed consistently across years. We had the 121 122 following final sample size in all years: 22 plots in October (except 2011) and 20 plots in 123 January (except 1990 and 2011) at PT; 20 plots in October (except 2011) and 20 plots in January (except 2001, 2011, 2012, and 2017) at CDB. Although we aim to survey the same stakes each 124 occasion, we sometimes cannot find stakes or miss a stake, so there is a slight mismatch in which 125 stakes were sampled on the October and January surveys. 126

We used October surveys to determine trends in active nests at the colony. We defined active nests as a nest with a penguin or egg present (Rebstock et al. 2016). Colony trends from October surveys are robust to observer variation and to variation in the number of stakes surveyed and survey start date (Rebstock et al. 2016). Active nests counted during the October survey are likely to overestimate the number of breeding pairs, as some of the males in active nests do not find mates.

To estimate OSR, we used January surveys, because some females and unmated males have not yet returned to the breeding colony by October (Boersma et al. 2013). We used a resampling approach to ensure that our sample size was sufficient to estimate trends in OSR accurately (Appendix Text S2). We calculated OSR as relative male abundance (No. Males/No. Males + No. Females) following Ancona et al. (2017). We train all field season volunteers to visually sex Magellanic penguins using cues including bill depth (males have noticeably deeper
bills), frons shape (males have steeper-sloped frons), body size (males are larger), and
aggressiveness (males tend to be more aggressive).

We estimated our visual sexing accuracy using known-aged banded birds that had been 141 sexed both visually and using at least one other method (n = 56,966 visual sight records). Other, 142 143 more certain, methods included: 1) genetic testing, 2) cloaca size relative to mate's cloaca size (Boersma and Davies 1987), 3) breeding behavior (copulation, etc.), 4) bill size (Bertellotti et al. 144 145 2002, Boersma, unpublished data), 5) sexual organs during necropsy of dead individuals. Our visual assessment of sex was correct between 86 and 100% of the time depending on the year, 146 suggesting low year-to-year variance in visual sexing accuracy. Year-specific accuracy was \geq 147 95% for females in all but two years (1984: 88% accuracy, 7 of 8 assessments; 2001: 50% 148 accuracy, 1 of 2 assessments) and for males in all but one year (1984: 93% accuracy, 13 of 14 149 150 assessments).

151 We did not adjust our estimate of OSR using sex-specific resignting rates (Ancona et al. 2017). At PT, resighting rates of females are lower than those of males, but resighting rates of 152 breeders of both sexes is high (male breeders: 0.97±0.02 SD, female breeders: 0.89±0.08 SD; 153 154 Gownaris and Boersma 2019). Many of the males counted during surveys are likely to be nonbreeders and resighting rates of non-breeders of either sex are substantially lower than those of 155 156 breeders (male non-breeders: 0.77±0.16 SD; female non-breeders: 0.47±0.21 SD; Gownaris and 157 Boersma 2019). Because males are more likely to be non-breeders than females (this study), 158 OSR may be more male-biased than our estimates, which assume equal detectability. 159 We used resighting records for banded birds to examine whether interannual variation in

160 OSR may have been influenced by survey date. Male Magellanic penguins settle at the colony in

September (Boersma et al. 2013). Females arrive approximately one week later and lay eggs in 161 October, then males generally make the first long incubation foraging trip (Boersma et al. 1990, 162 Yorio and Boersma 1994). Males and females trade off every day or two days when chicks are 163 small and require frequent feedings, then begin to make longer foraging trips, eventually leaving 164 large chicks in their nests alone. Therefore, OSR should be highly male-biased in September, 165 166 become female- then male- biased during the long incubation foraging trips, then stabilize during chick rearing. Chicks begin fledging in January, but peak fledging occurs in February (Boersma 167 168 et al 1990, Rebstock and Boersma 2017).

169

170 <u>Colony Monitoring: Breeding Performance and Male Fights at Punta Tombo</u>

In addition to annual surveys, we have intensively monitored PT throughout the breeding 171 season (September-February) every year since 1982, except for 2011. We search for banded 172 birds throughout the colony and on beaches. We mark and systematically monitor the nests of 173 174 banded birds, recording information on the behavior, health, and morphometrics of eggs, chicks, and adults. When we find a nest with eggs, we follow that nest daily until chicks hatch. We 175 measure the size of eggs at laying using calipers (length and width to the nearest mm) and 176 177 calculate egg volume (equations in Boersma and Rebstock 2010). We sum the volume of twoegg clutches to calculate clutch volume. Once all chicks hatch, we check most areas on a five to 178 179 ten-day cycle, except for one area where we check nests with chicks at approximately the same 180 time every day ("daily check area"). The daily check area covers ~0.7 ha and consists mainly of 181 bush nests (Rebstock and Boersma 2017). We band all adults found in a nest in the daily check 182 area. Because they are banded as adults, the age of these penguins is unknown.

183	We used data from banded females throughout the colony to calculate the following									
184	indicators of annual reproductive performance: egg-laying date, clutch volume, and reproductive									
185	success (Rebstock and Boersma 2018). To calculate median egg-laying date, we used nests for									
186	which we knew egg-laying date within two days, resulting in a sample size of 9-136 nests									
187	depending on year (75.5 \pm 31.1 SD nests per year). Our sample size for mean clutch volume was									
188	38-143 nests and only included two-egg nests (82.8 \pm 26.7 SD nests per year); nearly all									
189	Magellanic penguins lay two eggs and, if eggs or chicks are lost, they rarely re-lay eggs									
190	(Boersma et al. 2013). For reproductive success, we only used nests for which we knew the exact									
191	date that eggs were laid. We calculated reproductive success as the ratio of the number of chicks									
192	that fledged to the number of eggs laid (9-127 nests per year, average of 72.6 ± 29.3 SD). We									
193	assumed that chicks fledged if we saw them after January 10 weighing at least 1,800 grams									
194	(Boersma et al. 1990), as January 10 is the earliest we have seen a chick fledge since we started									
195	consistently recording fledging dates in 1988 (Boersma, unpublished data).									
196	We only included birds sighted in a nest in the daily check area in our calculations of sex-									
197	specific breeding probability. We also used data from this area of the colony to calculate a									
198	second estimate of OSR for each year. Data from the daily check area complement the survey									
199	data because, while they only represent a subset of the colony, we are certain of the breeding									
200	status and colony attendance of birds nesting there. Since we have sighted many of these birds									
201	regularly over several seasons, we are also certain of their sex in nearly all cases (2,521 of 2,675									
202	individuals nesting in this area since 1983). We categorized individuals as breeders if they were									
203	sighted with eggs or chicks. In some cases, unmated males at PT will attend nests with eggs or									

204 chicks that are not theirs; we did not consider these males to be breeders.

205	In 2000, we started consistently recording signs of penguin fights (blood, cuts on face or									
206	flippers), which most commonly occur between two males. Studies show that males fight for									
207	nests and for mates; fights are more likely to occur at high-quality nest sites and, among nests of									
208	ame quality, are more likely to occur at nest sites with a female present (Renison et al.									
209	02). Males that win fights are more likely to breed the following year than those that do not									
210	enison et al. 2002). Fighting occurs throughout the breeding season, but is most frequent									
211	uring settlement (i.e., in late September) and least frequent when most males are on their long									
212	incubation foraging trip (i.e., in late October) (Renison et al. 2006). Most often, males fight									
213	within nests (Renison et al. 2006), which could lead to the death of eggs or chicks.									
214	Fights are a relatively infrequent occurrence (Renison et al. 2006) and signs of fighting									
215	remain for five days or less, depending on how often the penguin goes to sea (Stor et al. 2019).									
216	Because our sample size for individuals with signs of fights was low ($n = 2-53$ across years, 385									
217	records total), we did not limit these data to the daily check area (i.e., we used sightings									
218	throughout the colony). We measured fight frequency as the number of banded males sighted at									
219	least once with signs of a fight for every 1000 unique males sighted that season.									
220	We used R Version 3.3.3 for all statistical analyses (R Core Development Team 2017).									
221	To test Hypothesis 1, we created a series of linear and mixed effects linear models to test for									
222	significance of year, colony, and stake number in driving OSR. We tested for spatial variation in									
223	OSR using the Akaike information criterion (AIC) of three models of varying complexity at each									
224	colony: a model containing only year (linear model), a model containing random intercepts									
225	across stakes (mixed effects model), and a model containing random slopes across stakes (mixed									
226	effects model) (package nlme, Version 3.1-131.1).									

For each colony, we ran a linear model testing for temporal trends in the number of active 227 nests and in survey-estimated OSR. Due to apparent non-linearity in the temporal trend of active 228 229 nests at CDB, we tested for breakpoints in these trends using mean squared error (Appendix Text S1). For survey-estimated OSR, we tested whether temporal trends of the two colonies from 230 1993-2016 were similar by using AIC to compare models that included: 1) year, 2) additive 231 232 terms for year and colony, or 3) interaction terms for year and colony. We used residuals from 233 colony-specific linear temporal trends in OSR for 1993-2016 to test for synchronicity in the 234 interannual variation in OSR across colonies. We tested for temporal autocorrelation in OSR 235 estimates using the acf function in the R stats package, using the longest contiguous stretch of data and considering auto-correlation over a 10-year period. 236

To test Hypothesis 2.1, we used colony-wide data on the three reproductive performance parameters of interest (egg-laying date, clutch volume, reproductive success) and surveyestimated OSR. When these parameters of interest showed temporal trends, we tested for relationships with variation in OSR by regressing the temporal residuals of the parameter on the temporal residuals for OSR. By using residuals, we tested how interannual variation in these parameters correlated with interannual variation in OSR rather than testing for similar or dissimilar long-term trends.

Hypothesis 2.2 and 2.3 focused on data collected in the daily check area at PT, as we are certain of the breeding status of penguins nesting in this area. We tested for temporal trends in the OSR of this area, in the breeding probability of males and females in this area, and in the frequency of male fights across the colony. We regressed residuals from temporal trends in the breeding probability and fight frequency parameters on residuals from temporal trends in daily check-estimated OSR.

250

271

251 **Results**

252 <u>Hypothesis 1: Trends and Variation in OSR</u>

In January 1993 to 2016, survey-estimated OSR varied from 0.43 (1998) to 0.82 (2010) 253 at PT and from 0.44 (1998) to 0.86 (2010) at CBD (Figure 2). Both colonies had a male-biased 254 OSR in all years except for 1998 (Figure 2). The mean OSR of the two colonies from 1993-2016 255 (PT: 0.68 ± 0.10 SD; CDB 0.68 ± 0.11 SD) was similar (t = 0.94, df = 39, p=0.35). We found no 256 temporal autocorrelation in survey-estimated OSR at either colony (PT: $r^2 < 0.07$ and p > 0.05257 for all autoregressive models out to 10 years; CDB: $r^2 < 0.11$ and p > 0.05 for all autoregressive 258 models out to 10 years). 259 Of the course of the study (1983-2017), survey-estimated OSR at PT became consistently 260 more male-biased (b = 0.0046, $r^2 = 0.20$, p = 0.01); the slope of this relationship steepened after 261 1992 (b = 0.0073, $r^2 = 0.35$, p < 0.005) (Figure 2). Similarly, survey-estimated OSR became 262 more male-biased from 1992-2016 at CDB (b = 0.0080, r² = 0.28, p < 0.01) (Figure 2). When 263 combining data from the two colonies for 1993 to 2016, a model containing only year as a 264 predictor of OSR (b = 0.0075, r² = 0.30, p < 0.0001) had similar support to a model that 265 266 contained different intercepts for PT and CDB ($\Delta AIC = 1.12$) and more support than a model that contained different slopes and intercepts for PT and CDB ($\Delta AIC = 3.10$). Interannual 267 268 variation in OSR, measured by the residuals of the temporal trends for each colony, was strongly correlated between PT and CBD (b = 0.98, $r^2 = 0.64$, p < 0.0001). 269 Interannual variation in the number of females counted during surveys was positively 270

272 (b = 0.50, r^2 = 0.40, p<0.001) and CBD (b = 0.37, r^2 = 0.49, p<0.0001). At both colonies,

correlated with interannual variation in the number of males counted during surveys at both PT

interannual variation in survey-estimated OSR was negatively correlated with interannual variation in the number of females counted (PT: $b = -3.5 \times 10^{-3}$, $r^2 = 0.36$, p <0.0005; CBD: b = -1.7x10⁻³, $r^2 = 0.22$, p=0.02) but was not correlated with variation in the number of males (PT: r^2 = 0.03, p=0.39; CBD: $r^2 = 0.06$, p=0.28).

Trends in OSR did not vary spatially at PT or at CDB (Figure 3). At PT, the mixed 277 278 effects model including random slopes (i.e., OSR trends varying with stake) was overfitted and the model including random intercepts did not perform as well as the model containing only the 279 fixed effect of year ($\Delta AIC = 20.24$). At CDB, a model containing only the fixed effect of year 280 281 had better support than a model containing random slopes ($\Delta AIC = 23.00$) or intercepts ($\Delta AIC =$ 20.00). The simplest and best-supported models at each colony both showed a significantly more 282 male-biased OSR over time, but had lower descriptive power than models aggregated at the year 283 level (PT: b = 0.0077, $r^2 = 0.07$, p < 0.0001; CBD: b = 0.0096, $r^2 = 0.09$, p < 0.0001). 284

The slope of daily check-estimated OSR at PT was shallower than survey-estimated OSR 285 at PT (b = 0.0055, $r^2 = 0.80$, p < 0.0001) and showed different patterns of interannual variation (r^2 286 = 0.002, p = 0.82). Daily check-estimated OSR showed positive temporal autocorrelation, with 287 an $r^2 = 0.67$ (p < 0.05) at one year that decreased to $r^2 = 0.19$ after five years (p < 0.05). The 288 289 number of females sighted in a nest in this area declined significantly from 1983-2017 (b = -2.8, $r^2 = 0.58$, p < 0.0001) but the number of males showed no temporal trend. Interannual variation in 290 the number of males nesting in the daily check area was correlated with variation in the number 291 of females nesting in this area (b = 0.60, $r^2 = 0.74$, p < 0.0001). As with the survey data, 292 interannual variation in daily check-estimated OSR was negatively correlated with variation in 293 the number of females (b = $-6x10^{-4}$, r² = 0.32, p < 0.001), but not with the number of males (r² = 294 295 0.03, p = 0.31).

Between 1983 and 2017, we collected over 340,000 sight records of males and over 296 220,000 sight records of females throughout PT. In Figure 2, we show seasonal variation in OSR 297 calculated using two approaches: 1) by aggregating all sight records at the day level across years 298 then calculating OSR (e.g., using the sum of males seen on October 10 and the sum of all 299 females seen on October 10 for all years to calculate OSR for that day) and 2) by calculating 300 301 OSR for each date of each year then averaging across years (e.g., the OSR for October 10 during each year averaged across 1987-2017). The two approaches resulted in similar patterns, though 302 303 sex-ratio was consistently more male-biased using the second approach. Because the second 304 approach is more volatile due to low sample sizes on some days (Figure 2), we focus our discussion on the results obtained using the first approach. As expected, OSR varied throughout 305 the season, approaching 1.00 at the start of the season (male settlement), and reaching values as 306 low as 0.22 during the last week of October, when most males are on their long incubation trip 307 (Figure 2). OSR was male-biased and relatively stable over the period when we conducted 308 January surveys (January 10 – February 7 depending on year), ranging from 0.65-0.70 (average 309 of 0.67 ± 0.015 SD) so it is unlikely that survey date impacted OSR variation across years. 310

311

312 Hypothesis 2: OSR and Breeding

Median egg-laying date in the daily check area at PT ranged from October 7 to October 22 and became progressively later from 1983-2017 (b=0.33, $r^2 = 0.57$, p<0.0001) (Figure S1). The residuals from the temporal regressions for egg-laying date were normally distributed and were not correlated with interannual variation in nest check-estimated OSR at PT ($r^2 = 0.01$, p = 0.55). Mean clutch volume in the daily check area at PT was 232.3 ± 5.1 and increased from 1983-2017 (b = 0.18, $r^2 = 0.14$, p = 0.04) (Figure S1). The temporal trend residuals were normally distributed and were not related to variation in OSR ($r^2 = 0.03$, p = 0.37). Reproductive success had high interannual variation (average 0.48 ± 0.26 SD) that was not related to surveyestimated OSR at PT ($r^2 = 0.01$, p = 0.50) and did not show a consistent temporal trend ($r^2 =$

322 6.8×10^{-4} , p = 0.88) (Figure S1).

Based on the estimated linear trend, the percentage of males breeding in the daily nest check area at PT decreased at a rate of approximately 1% per year from 1983-2017 (b = -1.1, r^2 = 0.56, p < 0.0001) (Figure 4). Male breeding rates in this area of the PT colony averaged 65% in the first three years of the study and only 29% in the most recent three years of the study (Figure 4). Over the same time period, the percentage of females breeding increased (b = 0.26, r^2 = 0.13, p = 0.03) (Figure 4). Female breeding rates in the daily nest check area of PT were never below 67% and were >80% in all but three years (1983, 1986, 1993).

330 Male fighting frequency increased from 2000-2017 (b = 2.44, r² = 0.64, p = 0.0001).

Fighting was more common in years when the sex ratio was more male-biased (b = 0.22, r^2 =

0.45, p = 0.003) (Figure 4). However, this trend was no longer significant with 2010 removed,

which represents an outlying year for both parameters ($r^2 = 0.20$, p = 0.09).

334

335 Discussion

In 1939, Mayr called for more research into sex ratios and their variability, but this topic
remains understudied, particularly in long-lived organisms (Mayr 1939, Székely et al. 2014a,
Carmona-Isunza et al. 2017). Though male-biased sex ratios are frequently reported in penguins
(yellow-eyed penguin *Megadyptes antipodes*: Richdale 1957; Adélie penguin *Pygoscelis adeliae*:
Ainley and Demaster 1980; King penguin *Aptenodytes patagonicus*: Olsson and van der Jeugd
2002; Galápagos penguin *Spheniscus mendiculus*: Boersma et al. 2013; southern rockhopper

342 penguin *Eudyptes chrysocome*: Morrison et al. 2015; African penguin *Spheniscus demersus*:

343 Spelt and Pichegru 2017), generally only static estimates of sex ratio are available for these

344 species (see Richdale 1957, Ainley and Demaster 1980, Weimerskirch et al. 2005, Gownaris and

Boersma 2019 for age-related trends). We leveraged a long-term dataset on two Magellanic

346 penguin colonies to examine how OSR varies at several temporal scales (long term,

interannually, intraannually) and to explore how this variation influences breeding behavior andsuccess in this long-lived, monogamous species.

349

350 <u>Hypothesis 1: Trends and Variation in OSR</u>

351 As predicted, OSR became more male-biased over time at both PT and CBD (Prediction 352 H1.1) and interannual variation in OSR at these colonies was synchronous (Prediction H1.2). At both colonies, interannual variation in OSR was related to variation in the number of females, 353 354 but not males, counted (Prediction H1.3). These similarities in patterns of OSR existed despite dissimilarities in the size, density, and habitat quality of the two colonies (Appendix Text S1). 355 Instead, similarities across the two colonies are likely to be driven by foraging conditions during 356 357 the non-breeding season, which previous studies suggest have sex-specific impacts on Magellanic penguins. As outlined in the introduction, females have lower non-breeding survival 358 359 rates than males, and this sex-bias in survival is greatest in juveniles (i.e., inexperienced 360 foragers) and in years when male survival is relatively low, an indication of poor foraging conditions in this food-limited species (Gownaris and Boersma 2019). Recapture rates also 361 suggest that females of this species are more likely to skip breeding than are males (Boersma and 362 363 Rebstock 2010) and reduced food during the non-breeding season availability (i.e., a stronger Rio de la Plata plume) influences the body condition of female but not male Magellanic penguins 364

(Rebstock and Boersma 2018). Links between sex-specific demography and non-breeding season
food availability have not been explicitly tested in this species and would substantially improve
our understanding of what drives the species' increasingly male-biased and variable OSR.

At PT, OSR trends and patterns in sex ratio varied with data source. Data from annual 368 surveys provided a snapshot of OSR throughout the colony, while data from daily nest checks 369 370 represented a restricted area of the colony throughout the breeding season. Although their total variance was similar, the long-term trend accounted for more than double the variance in nest 371 372 check-estimated OSR than in survey-estimated OSR. The limited temporal scale of survey-373 estimated OSR data may lead to greater stochasticity in these data, though sighting records throughout the colony did suggest that OSR is relatively stable at PT throughout the period that 374 surveys occurred. Temporal autocorrelation in OSR is expected for in long-lived species with 375 overlapping generations (Engen et al. 2003) but was only evident using nest-check data. 376 377 The appropriate method for tracking OSR will depend on the species and question of 378 interest. Data collected at high temporal frequency in smaller geographic areas may provide more robust estimates of temporal trends in OSR, but these data would not be appropriate in 379 cases where OSR varies spatially. For example, the presence of habitat refugia influences site-380 381 specific sex ratio in water striders (Aquarius remiges; Krupa & Sih 1993) and habitat-specific predation risk drives sex ratio variation in brown hares (Lepus europaeus; Husek et al. 2015), so 382 383 studying only subsets of these populations would lead to biased estimates of OSR. Because we 384 did not find that stakes throughout PT and CDB differed in their intercepts and slopes when 385 considering temporal trends in OSR, we found no evidence for strong spatial patterns in this

387

386

388 Hypothesis 2: OSR and Breeding

parameter at the two colonies.

Though we found evidence that interannual variation in OSR at Punta Tombo was linked 389 to the number of females at the colony, suggesting skipped breeding may be an underlying cause 390 391 of this variation, we found no link between OSR and reproductive success (Prediction H2.1). One possibility is that, following low food availability during the non-breeding season, the body 392 condition of returning females is effected (Rebstock and Boersma 2018), but that the females in 393 394 below a certain threshold do not return to breed. Females that return to breed in these years may also be capable of finding high quality mates due to low female-female competition or may face 395 396 lower intra-specific competition for food while foraging for chicks (e.g., Furness and Birkhead 397 1984, Lewis et al. 2001, Pozzi et al 2015).

One of the few previous studies examining spatiotemporal patterns in OSR in a seabird, the common eider (*Somateria mollissima*), found that skipped breeding was not correlated with OSR (Öst et al. 2018). However, reproductive success was lower in years when fewer females bred (Öst et al. 2018). In this system, predation risk is a significant cause of nest loss (19-22% of nests depredated) and spatiotemporal patterns in predation risk were related to female breeding decisions. In our system, about 40% of chicks starve and only 9% of chicks are lost to predators, so predation is not an important driver of chick death (Boersma and Rebstock 2014).

The absence of a correlation between reproductive success and OSR also suggests that aggression of unmated males at PT towards eggs and chicks does not significantly drive colonywide productivity in this species. However, we did find that fights between males, which can sometimes cause chick death, increased over time at PT. Fights were more common in years with more male-biased OSR, though this finding in influenced by an outlying year for both parameters (Prediction H2.2). This trend occurred even though male-male fights are more common in high-density areas of PT (Stokes and Boersma 2000) and the density of the colony

has declined over time (Rebstock et al. 2016). Though we do not have data on male harassment 412 of females at PT, heavily male-biased OSR can drive increased rates of female-harassment, 413 414 leading to a positive feedback loop that promotes further population decline (e.g., in common lizard Lacerta vivpara: Le Galliard et al. 2005; in common eider: Öst et al. 2018). 415 Nest quality benefits reproductive success in Magellanic penguins (Stokes and 416 417 Boermsa 1998), but this is unlikely to be the only motivation for a male penguin to fight. Nest sites are not a highly limited resources at PT. While the breeding population of the colony has 418 419 rapidly declined over the past 35 years, the perimeter of the colony has not changed (Rebstock et 420 al. 2016). Instead, the density of the colony has declined over this time, and even nests in highquality habitat may go unused each year (Rebstock et al. 2016). It is therefore likely that the 421 benefit to males comes, in part, from their ability to attract a female during the next breeding 422

male Magellanic penguin's chance of mating the following year, as divorce rates in some
seabirds (e.g., in short-tailed shearwater *Puffinus tenuirostris*: Bradley et al. 1990) and across

season (Renison et al. 2002). Aggression towards eggs or chicks may also increase an unmated

426 birds generally (Dubois and Cézilly 2002) are higher following failed breeding attempts.

423

427 Males should have a stronger incentive to fight when females are a limiting resource, 428 i.e., under more male-biased OSR (Renison et al. 2006). At PT, the probability of a male finding a mate declined by approximately 1% annually over the 34 years and was lower in years with 429 430 more male-biased OSR (Prediction H2.3), though the probability of females breeding did not 431 change over this time. Studies on other species of penguin have similarly shown that, over their 432 lifetime, the less abundant sex has more opportunities to mate than the more abundant sex (e.g., 433 yellow-eyed penguins: Richdale 1957). Because the total number of males sighted during daily 434 nest checks at PT showed no temporal trend, the absolute number of non-breeding males at the

435 colony increased over the study. In addition to affecting breeding behavior, the presence of non-436 breeders influences survival and fecundity rates of the breeding population (Lee et al. 2017).

At PT and CBD, long-term increases in OSR mean that effective population size is 437 declining more rapidly than suggested by trends in the number of active nests. The number of 438 males nesting in our daily check area has not declined over time, so population counts based on 439 440 occupied nests would suggest a stable population. However, trends in the number of females counted (2.8% annual decline) show that the effective population is declining. Similarly, in 441 442 snowy plovers (*Charadrius nivosus*), ignoring sex-specific survival rates leads to erroneously inflated population growth rates (Eberhart-Phillips et al. 2017). Given the direct impact of OSR 443 on the size of the effective breeding population and the impact that non-breeders can have on 444 breeders, accurately estimating and predicting trends in effective population size requires the 445 inclusion of sex ratio in monitoring studies. 446

447

448 Acknowledgements

The Center for Ecosystem Sentinels is located in the Biology Department at the 449 University of Washington. The Province of Chubut and the Offices of Tourism and of Flora and 450 451 Fauna provide permits and living space during the field season. We are grateful to the La Regina family for allowing us to work on their land. The Wildlife Conservation Society (WCS) helped 452 453 fund the project and secure permits. Graham Harris and Pat Harris gave logistical support. All 454 research was conducted under approval of the University of Washington IACUC (Protocol 455 #2213-02). Financial support for fieldwork was provided by WCS, Exxonmobil Foundation, the 456 Pew Fellows Program in Marine Conservation, the Disney Worldwide Conservation Fund, the 457 Chase, Cunningham, CGMK, Offield, Peach, Thorne, Tortuga, and Kellogg Foundations. The

458	Wadsworth Endowed Chair in Conservation Science and Friends of the Penguins supported data
459	analysis. We thank C. Gravelle and O. Woods for database help and field volunteers and students
460	for help with data collection.
461	
462	Literature Cited
463	Ainley, D. G., and D. P. DeMaster. 1980. Survival and mortality in a population of Adélie
464	penguins. Ecology 61 :522-530.
465	Ancona, S., F. V. Dénes, O. Krüger, T. Székely, and S. R. Beissinger. 2017. Estimating adult sex
466	ratios in nature. Philosophical Transactions of the Royal Society B: Biological Sciences

467 372:20160313.

- 468 Bertellotti, M., J. L. Tella, J. A. Godoy, G. Blanco, M. G. Forero, J. A. Donázar, and O.
- 469 Ceballos. 2002. Determining sex of Magellanic penguins using molecular procedures and
 470 discriminant functions. Waterbirds 25:479-485.
- Boersma, P. D. and E. M. Davies. 1987. Sexing monomorphic birds by vent measurements. The
 Auk 104:779-783.
- 473 Boersma, P. D., and G. A. Rebstock. 2010. Effects of double bands on Magellanic penguins.
 474 Journal of Field Ornithology 81:195-205.
- Boersma, P. D., and G. A. Rebstock. 2014. Climate change increases reproductive failure in
 Magellanic penguins. PLOS ONE 9:e85602.
- Boersma, P. D., G. A. Rebstock, E. Frere, and S. E. Moore. 2009. Following the fish: penguins
 and productivity in the South Atlantic. Ecological Monographs **79**:59-76.
- 479 Boersma, P.D., Frere, E., Kane, O., Pozzi, L.M., Pütz, K., Andrew, R.R., Rebstock, G.A.,
- 480 Simeone, A., Smith, J., Van Buren, A., Yorio, P., and García-Borboroglu, P. 2013.

481	Magellanic penguin. In P. García-Borboroglu and P.D. Boersma (Eds.), Penguins: natural
482	history and conservation (pp. 233-262). Seattle, WA: University of Washington Press.
483	Boersma, P.D., Stokes, D.L., and P.M. Yorio. 1990. Reproductive variability and historical
484	change of Magellanic penguins (Spheniscus magellanicus) at Punta Tombo, Argentina. In
485	L.S. Davis and J.T. Darby (Eds.), Penguin biology (pp. 15-41). San Diego, CA:
486	Academic Press.
487	Bradley, J.S., Wooller, R.D., Skira, I.J., and D.L. Serventy. 1990. The influence of mate
488	retention and divorce upon reproductive success in short-tailed shearwaters Puffinus
489	tenuirostris. Journal of Animal Ecology 59: 487-496.
490	Carmona-Isunza, M. C., S. Ancona, T. Székely, A. P. Ramallo-González, M. Cruz-López, M. A.
491	Serrano-Meneses, and C. Küpper. 2017. Adult sex ratio and operational sex ratio exhibit
492	different temporal dynamics in the wild. Behavioral Ecology 28:523-532.
493	Cubaynes, S., P. F. Doherty, E. Schreiber, and O. Gimenez. 2011. To breed or not to breed: a
494	seabird's response to extreme climatic events. Biology Letters 7:303-306.
495	Donald, P. F. 2007. Adult sex ratios in wild bird populations. Ibis 149:671-692.
496	Drent, R.H. and S. Daan. 1980. The prudent parent: adjustments in avian breeding. Ardea 68:
497	225-252.
498	Dubois, F. and F. Cézilly. 2002. Breeding success and mate retention in birds: a meta-analysis.
499	Behavioral Ecology and Sociobiology 52: 357-364.
500	Eberhart-Phillips, L. J., C. Küpper, M. C. Carmona-Isunza, O. Vincze, S. Zefania, M. Cruz-
501	López, A. Kosztolányi, T. E. Miller, Z. Barta, and I. C. Cuthill. 2018. Demographic
502	causes of adult sex ratio variation and their consequences for parental cooperation.
503	Nature Communications 9:1651.

504	Eberhart-Phillips, L. J., C. Küpper, T. E. Miller, M. Cruz-López, K. H. Maher, N. Dos
505	Remedios, M. A. Stoffel, J. I. Hoffman, O. Krüger, and T. Székely. 2017. Sex-specific
506	early survival drives adult sex ratio bias in snowy plovers and impacts mating system and
507	population growth. Proceedings of the National Academy of Sciences 114:E5474-E5481.
508	Engen, S., R. Lande, and BE. SÆther. 2003. Demographic stochasticity and Allee effects in
509	populations with two sexes. Ecology 84:2378-2386.
510	Ewen, J. G., R. Thorogood, and D. P. Armstrong. 2011. Demographic consequences of adult sex
511	ratio in a reintroduced hihi population. Journal of Animal Ecology 80:448-455.
512	Furness, R. W. and T. R. Birkhead. 1984. Seabird colony distributions suggest competition for
513	food supplies during the breeding season. Nature 311: 655-656.
514	Fisher, R. A. 1930. The genetical theory of natural selection: a complete variorum edition.
515	Oxford, UK: Oxford University Press.
516	Gownaris, N. J. and P.D. Boersma. 2019. Sex-biased survival contributes to population decline
517	in a long-lived seabird, the Magellanic penguin. Ecological Applications 29:e01826.
518	Heinsohn, R., G. Olah, M. Webb, R. Peakall, and D. Stojanovic. 2019. Sex ratio bias and shared
519	paternity reduce individual fitness and population viability in a critically endangered
520	parrot. Journal of Animal Ecology 88:502-510.
521	Hušek, J., M. Panek, and P. Tryjanowski. 2015. Predation Risk Drives Habitat-Specific Sex
522	Ratio in a Monomorphic Species, the Brown Hare (Lepus europaeus). Ethology 121:
523	593-600.
524	Jenouvrier, S., H. Caswell, C. Barbraud, and H. Weimerskirch. 2010. Mating behavior,
525	population growth, and the operational sex ratio: a periodic two-sex model approach. The

526 American Naturalist **175**:739-752.

527	Krupa, J. J. and A. Sih. 1993. Experimental studies on water strider mating dynamics: spatial
528	variation in density and sex ratio. Behavioral Ecology and Sociobiology 33: 107-120.
529	Lambertucci, S. A., M., Carrete, J. A., Donázar and F. Hiraldo. 2012. Large-scale age-dependent
530	skewed sex ratio in a sexually dimorphic avian scavenger. PLOS ONE 7:e46347.
531	Le Galliard, JF., P. S. Fitze, R. Ferrière, and J. Clobert. 2005. Sex ratio bias, male aggression,
532	and population collapse in lizards. Proceedings of the National Academy of Sciences
533	102 :18231-18236.
534	Lee, A. M., J. M. Reid, and S. R. Beissinger. 2017. Modelling effects of non-breeders on
535	population growth estimates. Journal of Animal Ecology 86:75-87.
536	Lewis, S., T. N. Sherratt, K. C. Hamer, and S. Wanless. 2001. Evidence of intra-specific
537	competition for food in a pelagic seabird. Nature 412 : 816-819.
538	Liker, A., T. Székely, and T. Tregenza. 2005. Mortality costs of sexual selection and parental
539	care in natural populations of birds. Evolution 59 :890-897.
540	Mayr, E. 1939. The sex ratio in wild birds. The American Naturalist 73 :156-179.
541	Morrison, C. A., R. A. Robinson, J. A. Clark, and J. A. Gill. 2016. Causes and consequences of
542	spatial variation in sex ratios in a declining bird species. Journal of Animal Ecology
543	85 :1298-1306.
544	Morrison, K. W., P. F. Battley, P. M. Sagar, and D. R. Thompson. 2015. Population dynamics of
545	Eastern Rockhopper Penguins on Campbell Island in relation to sea surface temperature
546	1942–2012: current warming hiatus pauses a long-term decline. Polar Biology 38:163-
547	177.
548	Olsson, O., and H. P. van der Jeugd. 2002. Survival in king penguins Aptenodytes patagonicus:
549	temporal and sex-specific effects of environmental variability. Oecologia 132:509-516.

550	Öst, M., A. Lindén, P. Karell, S. Ramula, and M. Kilpi. 2018. To breed or not to breed: drivers
551	of intermittent breeding in a seabird under increasing predation risk and male bias.
552	Oecologia 188 :129-138.
553	Pöysä, H., P. Linkola, and A. Paasivaara. 2019. Breeding sex ratios in two declining diving duck
554	species: between-year variation and changes over six decades. Journal of Ornithology

- 555 doi.org/10.1007/s10336-019-01682-7.
- Pozzi, L. M., P. G. Borboroglu, P. D. Boersma, and M. A. Pascual. 2015. Population regulation
 in Magellanic penguins: what determines changes in colony size? PLOS ONE
 10:e0119002.
- R Core Development Team (2017). R: A language and environment for statistical computing. R
 Foundation for Statistical Computing, Vienna, Austria. URL <u>http://www.R-project.org/.</u>
- Rankin, D. J., U. Dieckmann, and H. Kokko. 2011. Sexual conflict and the tragedy of the
 commons. The American Naturalist 177: 780-791.
- Rebstock, G. A., P. D. Boersma, and P. García-Borboroglu. 2016. Changes in habitat use and
 nesting density in a declining seabird colony. Population Ecology 58:105-119.
- 565 Rebstock, G. A. and P. D. Boersma. 2017. Comparing reproductive success of a colonial seabird,
- the Magellanic penguin, estimated by coarse- and fine-scale temporal sampling. The
- 567 Condor: Ornithological Applications **119**:225-238.
- 568 Rebstock, G. A., and P. D. Boersma. 2018. Oceanographic conditions in wintering grounds
- affect arrival date and body condition in breeding female Magellanic penguins. Marine
- 570 Ecology Progress Series **601**:253-267.

571	Reed, T. E., M. P. Harris, and S. Wanless. 2015. Skipped breeding in common guillemots in a
572	changing climate: restraint or constraint? Frontiers in Ecology and Evolution
573	https://doi.org/10.3389/fevo.2015.00001.

- 574 Renison, D., D. Boersma, D., and M. B. Martella. 2002. Winning and losing: causes for
- variability in outcome of fights in male Magellanic penguins (*Spheniscus magellanicus*).
 Behavioral Ecology 13:462-466.
- Renison, D., P. D. Boersma, A. N. Van Buren, and M. B. Martella. 2006. Agonistic behavior in
 wild male Magellanic penguins: when and how do they interact? Journal of Ethology
 24:189-193.
- 580 Richdale, L. E. 1957. A Population Study of Penguins. Oxford, UK: Oxford University Press.
- Salton, M., C. Saraux, P. Dann, and A. Chiaradia. 2015. Carry-over body mass effect from
 winter to breeding in a resident seabird, the little penguin. Royal Society Open Science
 2:140390.
- Saino, N., R. Ambrosini, M. Caprioli, A. Romano, M. Romano, D. Rubolini, C. Scandolara, and
 F. Liechti. 2017. Sex-dependent carry-over effects on timing of reproduction and

fecundity of a migratory bird. Journal of Animal Ecology **86**:239-249.

- 587 Schreiber, E. A., and J. Burger. 2001. Biology of marine birds. Boca Raton, FL: CRC Press.
- Spelt, A., and L. Pichegru. 2017. Sex allocation and sex-specific parental investment in an
 endangered seabird. Ibis 159:272-284.
- Stokes, D. L., and P. D. Boersma. 1998. Nest-site characteristics and reproductive success in
 Magellanic penguins (*Spheniscus magellanicus*). The Auk 115:34-49.
- Stokes, D. L., and P. D. Boersma. 2000. Nesting density and reproductive success in a colonial
 seabird, the Magellanic penguin. Ecology 81:2878-2891.

- Stokes, D. L., P. D. Boersma, J. L. de Casenave, and P. García-Borboroglu. 2014. Conservation
 of migratory Magellanic penguins requires marine zoning. Biological Conservation
 170:151-161.
- 597 Stor, T., G. A. Rebstock, P. G. Borboroglu, and P. D. Boersma. 2019. Lateralization
 598 (handedness) in Magellanic penguins. PeerJ 7:e6936.
- Székely, T., F. Weissing, and J. Komdeur. 2014a. Adult sex ratio variation: implications for
 breeding system evolution. Journal of Evolutionary Biology 27:1500-1512.
- 601 Székely, T., A. Liker, R. P. Freckleton, C. Fichtel, and P. M. Kappeler. 2014b. Sex-biased
- survival predicts adult sex ratio variation in wild birds. Proceedings of the Royal Society
 B: Biological Sciences 281:20140342.
- Szekely, T., G. H. Thomas, and I. C. Cuthill. 2006. Sexual conflict, ecology, and breeding
 systems in shorebirds. BioScience 56:801-808.
- 606 Vanstreels, R. E. T., A. C. Adornes, P. L. Canabarro, V. Ruoppolo, M. Amaku, R. P. da Silva-
- Filho, and J. L. Catão-Dias. 2013. Female-biased mortality of Magellanic penguins
 (*Spheniscus magellanicus*) on the wintering grounds. Emu **113**:128-134.
- 609 Vanstreels, R. E. T., A. C. Adornes, V. Ruoppolo, P. L. Canabarro, R. P. Silva-Filho, and J. L.
- 610 Catao-Dias. 2011. Gender determination from morphometrics in migrating Magellanic
 611 penguins (*Spheniscus magellanicus*). Marine Ornithology **39**:215-220.
- 612 Weimerskirch, H., J. Lallemand, and J. Martin. 2005. Population sex ratio variation in a
- monogamous long-lived bird, the wandering albatross. Journal of Animal Ecology 74:285-291.
- Yorio, P., and P.D. Boersma. 1994. Causes of nest desertion during incubation in the Magellanic
 penguin (*Spheniscus magellanicus*). The Condor **96**:1076-1083.
- 617

618 Figure Legends

619

Figure 1: The Magellanic penguin breeding colonies of Punta Tombo (PT) and Cabo Dos Bahias 620 621 (CDB) are within 105 km of each other, as the crow flies, in the Province of Chubut, Argentina. The population at PT is an order of magnitude larger than that of CDB. Map insets show the 622 623 locations of plots we surveyed annually each October, to estimate temporal trends in population size. While PT's population trend has been consistently negative since 1987, the population trend 624 at CDB was positive until 2009, and since 2010 has shown a non-significant decline. Base map is 625 the "World Imagery" map provided by Esri (Sources: Esri, DigitalGlobe, Earthstar Geographics, 626 CNES/Airbus DS, GeoEye, USDA FSA, USGS, Aerogrid, IGN, IGP, and the GIS User 627 Community). 628 629 630 Figure 2: Variability in operational sex ratio (OSR) at Punta Tombo (PT), Argentina and Cabo Dos Bahias (CDB), Argentina across years (2a) and at Punta Tombo, Argentina within a 631 breeding season (2b). The OSR at both PT (b = 0.0046, $r^2 = 0.20$, p = 0.01) and CDB (b = 0.0046, $r^2 = 0.20$, p = 0.01) 632 0.0080, $r^2 = 0.28$, p < 0.01) has become more male-biased over time, with a similar trend at both 633 634 colonies between 1993-2016. Interannual variation at the two colonies (solid line), measured as

the residuals of the temporal trend, are highly correlated (b = 0.98, $r^2 = 0.64$, p < 0.0001). At PT,

636 sightings of banded birds suggest substantial intraannual variation in OSR, driven by sex-specific

timing of settlement and foraging trips (2b). This variation was estimated using two approaches:

- 638 1) by aggregating all sight records at the month and month day level across years then
- 639 calculating OSR (green line), and 2) by calculating OSR for each date of each year then
- averaging across years (blue line \pm 95% CI). The grey box represents the period during which
- 641 OSR surveys were conducted at the colony (January 10 February 7) across years.

Figure 3: Stake-specific temporal trends in operational sex ratio (OSR) at Punta Tombo,

Argentina and Cabo Dos Bahias, Argentina. OSR became more male biased across all stakes at both colonies, as shown by the distribution of values across stakes (3a) and the stake-specific slope in operational sex ratio over time (3b). Linear mixed effects models showed that, within a colony, stake did not have a significant impact on the intercept or slope of ASR, i.e., models including random slopes and intercepts in addition to the fixed effect "year" had less support than those including only the fixed effect "year". This finding suggests little spatial variation in OSR and OSR trends at the two colonies.

650

Figure 4: Temporal trends in male fight frequency throughout the colony of Punta Tombo (PT), 651 652 Argentina (males with signs of fighting / 1000 males) (4a) and male breeding probability (4c) in an area of PT checked daily. Male fighting became more frequent over time (b = 2.44, $r^2 = 0.64$, 653 654 p = 0.0001) and interannual variation in this parameter was correlated with the operational sex 655 ratio (OSR), with fights occurring more frequently in years with a more male-biased OSR (b =0.22, $r^2 = 0.45$, p = 0.003) (4b). Males at PT fight throughout the breeding season for high 656 quality nest sites and mates and signs of a fight (bloodied/cut face or flippers) last for several 657 658 days (4d). Male fights may be increasingly common due to fiercer competition for females, as the percentage of males that find a mate has declined over time (b = -1.1, $r^2 = 0.56$, p < 0.0001) 659 (4c). 660

662 Figure 1









678 Sex ratio is variable and increasingly male-biased at two colonies of Magellanic penguins
 679

680 Gownaris, N.J., García Borboroglu, P., Boersma, P.D.

681

682 Appendix 1: Additional Methodological Details and Colony Characteristics

683

684 Text S1: Colony Trends and Characteristics

The Magellanic breeding colony at PT (44°02'S, 65°02'S) has occupied approximately 400 ha since at least 1970 and consists of approximately 200,000 breeding pairs (Rebstock et al. 2016). The colony was established in the 1920s and reached its peak size in the 1960s-1970s (Boersma et al. 1990). It was the largest breeding colony of the species but is now in decline. The number of active nests at the colony decreased by 43% from 1987-2017 (b = -2.81, r² = 0.67, p<0.0001; Figure 1).

691

CDB (44°54'S, 65°32'S) is less than 20 ha in size and is located approximately 100 km south of PT (Capurro et al. 1988). The age of this colony is unknown and the colony has declined since Capurro et al (1988) surveyed the colony. The breeding population was estimated to be 12,000 pairs in 2010 (Pozzi et al. 2015). At CDB, the number of active nests increased by 41% from 1993-2009 (b = 4.4.1, r^2 =0.51, p < 0.001) then began to decline, but the decreasing trend for 2010-2017 was not significant (Figure 1).

698

699 Nest density was higher at CDB than at PT in all years that we completed October surveys at

both colonies (1992-2010; 2013-2017). We calculated nest density based on all potential nesting

701 sites, i.e. a nest cup was present, but a bird did not have to be. Nest density ranged from 13 (2017) to 24 (2005) nests per 100 m² at CDB and from 9 (2016) to 13 (1996) nests per 100 m² at 702 PT. Nests at PT had significantly higher cover than at CDB in 2008-2010, 2013-2015, and in 703 2017 (Table 1). The most frequent nest cover was <60% cover (category 4) at CDB and 60-79% 704 cover at PT (category 3). This disparity may be due in part to nest type. Burrow nests have 705 706 higher cover than do bush nests on average (Stokes and Boersma 1998, Rebstock et al. 2016) and burrow nests make up a higher percentage of nests counted during surveys at PT ($53.0 \pm 4.3\%$ 707 708 across years) than at CDB (28.2 \pm 5.9% across years).

709

710 Text S2: Survey Plot Number Sensitivity

To determine how sensitive our OSR calculations were to the number of survey plots, we randomly sampled between 1 and 18 survey plots 100 times each, running a linear model and extracting slope and fit parameters for each run. To summarize these data, we averaged the slope across all runs for a given number of survey plots and calculated 95% confidence intervals and the percent of linear relationships that were significant. For 19 and 20 survey plots, we collected these data for the maximum number of unique combinations (20 and 1, respectively).

717

At PT, the average slope for temporal change in OSR from 1986-2017 was between 0.0048 and 0.0059 regardless of the number of survey plots examined. The 95% confidence intervals of the slopes from these runs suggested that the "true" change in OSR, i.e. that measured using 20 survey plots, could be captured by surveying seven or more plots. However, at least 15 plots had to be surveyed for the slope to be significant at least 95% of the time. At C2B, the slope for change in OSR from 1993-2016 ranged from 0.0077-0.0093 depending on the number of plots

724	surveyed. The confidence intervals for the slope included the slope determined from 20 survey								
725	plots (0.0078) if at least nine plots were surveyed. At least 12 plots had to be surveyed for the								
726	temporal change in OSR to be significant at least 95% of the time. Our results show that twenty								
727	survey plots were sufficient to capture OSR trends at PT and CDB.								
728									
729	References								
730	Capurro, A., E. Frere, M. Gandini, P. Gandini, T. Holik, V. Lichtschein, and P. D. Boersma.								
731	1988. Nest density and population size of Magellanic penguins (Spheniscus								
732	magellanicus) at Cabo Dos Bahias, Argentina. The Auk 105:585-588.								
733	Boersma, P.D., Stokes, D.L., and P.M. Yorio. 1990. Reproductive variability and historical								
734	change of Magellanic penguins (Spheniscus magellanicus) at Punta Tombo, Argentina. In								
735	L.S. Davis and J.T. Darby (Eds.), Penguin biology (pp. 15-41). San Diego, CA:								
736	Academic Press.								
737	Pozzi, L. M., P. G. Borboroglu, P. D. Boersma, and M. A. Pascual. 2015. Population regulation								
738	in Magellanic penguins: what determines changes in colony size? PLOS ONE								
739	10 :e0119002.								
740	Rebstock, G. A., P. D. Boersma, and P. García-Borboroglu. 2016. Changes in habitat use and								
741	nesting density in a declining seabird colony. Population Ecology 58:105-119.								
742	Stokes, D. L., and P. D. Boersma. 1998. Nest-site characteristics and reproductive success in								
743	Magellanic penguins (Spheniscus magellanicus). The Auk 115:34-49.								
744									
745									
746 747									

748	Table S1:	Number an	d Density	of Magellanic	penguin	Nests at	Punta	Tombo an	d Cabo	Dos
-----	-----------	-----------	-----------	---------------	---------	----------	-------	----------	--------	-----

Bahias, Argentina

	Pui	nta Tombo	Cabo Dos Bahias			
Year	No. Total Nests	Density (nests/100 m ²)	No. Total Nests	Density (nests/100 m ²)		
1987	301	15.05	NA	NA		
1988	258	12.90	NA	NA		
1989	255	12.75	NA	NA		
1990	237	11.85	NA	NA		
1991	195	9.75	NA	NA		
1992	214	10.70	359	17.95		
1993	225	11.25	373	18.65		
1994	209	10.45	338	16.90		
1995	212	10.60	342	17.10		
1996	253	12.65	348	17.40		
1997	217	10.85	339	16.95		
1998	212	10.60	338	16.90		
1999	202	10.10	423	21.15		
2000	215	10.75	383	19.15		
2001	191	9.55	394	19.70		
2002	185	9.25	481	24.05		
2003	184	9.20	400	20.00		
2004	182	9.10	396	19.80		
2005	185	9.25	409	20.45		
2006	183	9.15	381	19.05		
2007	169	8.45	386	19.30		
2008	185	9.25	441	22.05		
2009	224	11.20	436	21.80		
2010	220	11.00	435	21.75		
2011	NA	NA	NA	NA		
2012	245	12.25	NA	NA		
2013	184	9.20	371	18.55		
2014	197	9.85	330	16.50		
2015	217	10.85	356	17.80		
2016	169	8.45	294	14.70		
2017	188	9.40	263	13.15		

	Punta Tombo Nest Quality			Cabo Dos Bahias Nest Quality			Quality Comparison		
Year	Average	Mode	Percent Burrow	Average	Mode	Percent Burrow	t	df	р
2007	3.04	4	58.6%	3.03	4	29.0%	-0.1	420	8.9E-01
2008	2.85	3	62.6%	3.09	4	31.7%	2.7	428	6.4E-03
2009	2.85	2	53.5%	4.23	4	42.0%	14.1	273	2.2E-16
2010	3.19	3	51.3%	3.52	4	28.8%	4.7	485	3.4E-06
2011	NA	NA	NA	NA	NA	NA	NA	NA	NA
2012	2.69	3	51.1%	NA	NA	NA	NA	NA	NA
2013	2.78	4	49.8%	3.18	4	24.4%	3.7	447	2.8E-04
2014	2.58	2	48.8%	3.03	2	25.6%	4.4	483	1.5E-05
2015	2.84	4	50.6%	3.21	4	25.4%	3.8	441	1.5E-04
2016	2.58	2	50.8%	2.70	2	23.0%	1.1	396	2.8E-01
2017	2.92	3	52.7%	3.26	4	23.6%	3.4	461	7.0E-04

Table S2: Quality of Magellanic penguin Nests at Punta Tombo and Cabo Dos Bahias, Argentina

Sex ratio is variable and increasingly male-biased at two colonies of Magellanic penguins

Gownaris, N.J., García Borboroglu, P., Boersma, P.D.

- **Appendix 2: Supplementary Figure**

Figure S1



Figure S1: Temporal trends in three breeding performance indicators at Punta Tombo, Argentina for 1982-2016. Median egg lay date, measured as the date in October, became later at the colony over the course of the study (b=0.33, $r^2 = 0.57$, p<0.0001). Average clutch volume, measured for two-egg clutches, increased over time (b = 0.18, r² = 0.14, p = 0.04). Reproductive success

- showed high interannual variability and no temporal trend ($r^2 = 6.8 \times 10^{-4}$, p = 0.88).